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1 **Responses of nitrogen concentrations and pools to multiple environmental**
2 **change drivers: a meta-analysis across terrestrial ecosystems**

3

4 **Running head:** Terrestrial N response to environmental change

5

6 **Abstract**

7 **Aim:** We sought to understand how the individual and combined effects of multiple environmental
8 change drivers differentially influence terrestrial nitrogen (N) concentrations and N pools and
9 whether the interactive effects of these drivers are mainly antagonistic, synergistic or additive.

10 **Location:** Worldwide.

11 **Time period:** Contemporary.

12 **Major taxa studied:** Plant, soil, and soil microbes in terrestrial ecosystems.

13 **Methods:** We synthesized data from manipulative field studies from 760 published articles to
14 estimate the individual, combined, and interactive effects of key environmental change drivers
15 (elevated CO₂, warming, N addition, phosphorus addition, increased rainfall, and drought) on plant,
16 soil and soil microbe N concentrations and pools using meta-analyses. We assessed the influences
17 of moderator variables on these effects through structural equation modeling.

18 **Results:** We found that (1) N concentrations and N pools were significantly affected by the
19 individual and combined effects of multiple drivers, with N addition (either alone or in
20 combination with another driver) showing the strongest positive effects; (2) the individual and
21 combined effects of these drivers differed significantly between N concentrations and N pools in
22 plants, but seldom in soils and microbes; (3) additive effects of driver pairs on N concentrations
23 and pools were much more common than synergistic or antagonistic effects across plants, soils,

24 and microbes; and (4) environmental and experimental factors were important moderators of the
25 individual, combined, and interactive effects of these drivers on terrestrial N.

26 **Main conclusions:** Our results indicate that terrestrial N concentrations and N pools, especially
27 those of plants, can be significantly affected by the individual and combined effects of
28 environmental change drivers, with the interactive effects of these drivers being mostly additive.
29 Our findings are important because they contribute to the development of models to better predict
30 how altered N availability affects ecosystem carbon cycling under future environmental changes.

31

32 **Keywords:** altered rainfall, combined effects, elevated CO₂, individual effects, interactive effects,
33 nitrogen addition, phosphorus addition, warming

34

35 **1 Introduction**

36 Human-induced changes in the biogeochemical cycling of key nutrient elements such as nitrogen
37 (N) can strongly influence ecosystem structure and function (Gruber & Galloway, 2008). Changes
38 in N availability can remarkably affect multiple biological processes, including plant
39 photosynthesis, symbiotic N fixation, and microbial N mineralization, with potential consequences
40 for terrestrial carbon (C) cycling and storage (Peñuelas et al., 2013; Vitousek, 2004; Yue et al.,
41 2016). Enhanced N availability across terrestrial ecosystems has been occurring with simultaneous
42 alterations in other environmental change drivers, including elevated atmospheric CO₂ (eCO₂),
43 warming, increased rainfall (rainfall⁺), and drought, which can have divergent effects (Galloway,
44 2005; Peñuelas et al., 2013; Sardans et al., 2017). For example, recent meta-analyses reported that
45 N enrichment can significantly increase plant and soil N concentrations (Lu et al., 2011), whereas
46 eCO₂ (Deng et al., 2015) and drought (He & Dijkstra, 2014) can have the opposite effect on plant

47 N concentrations. Despite our growing understanding of the individual effects that different
48 environmental change drivers may have on terrestrial N concentrations, few attempts have been
49 made to address whether and how such individual effects on N concentrations (percent of dry mass)
50 and N pools (absolute mass per unit area) may differ significantly across different terrestrial
51 compartments (i.e., plants, soils, and soil microbes). How these environmental change drivers
52 interact to affect N concentrations and N pools in different terrestrial ecosystem compartments also
53 remains unclear.

54 Changes in N concentrations and N pools in response to environmental change drivers are
55 likely to vary strongly among (and within) plants, soils, and microbial biomass (Lu et al., 2011;
56 Sardans et al., 2017). For example, the negative effect of eCO₂ on plant N concentration (Cotrufo,
57 Ineson, & Scott, 1998) does not necessarily indicate a reduction in the total amount of N in plants
58 (i.e., the plant N pool) because eCO₂ could simultaneously stimulate plant biomass production
59 (Nowak, Ellsworth, & Smith, 2004). Likewise, the negative effects of drought on plant N
60 concentration can be reversed when plant biomass proportionally decreases more than plant N
61 uptake under drought stress (He & Dijkstra, 2014). Human-induced phosphorus (P) fertilization
62 also significantly influences plant N concentration (Yuan & Chen, 2015), while N-P imbalances
63 following repeated P fertilization can alter the functions of both natural and managed ecosystems
64 (Peñuelas et al., 2013). These findings suggest the need to disentangle plant N concentration and
65 N pool responses from different environmental change drivers. Soil or microbial N concentrations
66 and pools also vary in response to environmental change drivers. For example, it has been shown
67 that N fertilization can significantly influence soil bulk density by mediating the activities of soil
68 fauna and microbes (Crill, Martikainen, Nykanen, & Silvola, 1994), and that the application of N
69 alone or in combination with other nutrients (e.g., P) can affect soil N pools differently (Fornara,

70 Banin, & Crawley, 2013). Although previous synthesis studies (Bai et al., 2013; Lu et al., 2011)
71 have addressed the potential responses of terrestrial N pools to multiple drivers, these studies have
72 not distinguished between N concentrations and N pools and have mainly focused on the effects
73 of individual drivers.

74 Multiple environmental change drivers likely act simultaneously and influence a wide range
75 of ecological and biogeochemical processes (Reich et al., 2006; Yue, Fornara, Yang, Peng, Peng,
76 et al., 2017); thus, the combined effects of multiple drivers on N cycling may be more important
77 than the corresponding individual effects. For instance, a recent meta-analysis (Li, Niu, & Yu, 2016)
78 showed how the combined effect of N and P additions on plant N concentration tends to be higher
79 than their individual effects. On the other hand, the positive effects of eCO₂ or N addition on plant
80 N concentration may be suppressed by the presence of another driver, such as drought (Delgado-
81 Baquerizo et al., 2013). Additionally, reductions in soil N concentrations under eCO₂ can be
82 nullified by simultaneous warming (Hovenden et al., 2008). Although warming can significantly
83 increase both plant and soil N pools (Bai et al., 2013), it can also promote drought stress, which
84 makes predictions of the N pool response to warming more complex and uncertain.

85 Another important knowledge gap is whether the interactive effects of multiple environmental
86 change drivers on terrestrial N concentrations and pools are additive or non-additive. Additive
87 interactions occur when the combined effect of two or more drivers is equal to or not significantly
88 different from the sum of the individual effects; otherwise, the interaction is either synergistic or
89 antagonistic (Jessica Gurevitch, Morrison, & Hedges, 2000; Zhou et al., 2016). Both additive and
90 nonadditive effects have been found in empirical and synthesis studies. For example, the
91 interactive effects of eCO₂ and warming on plant biomass were found to be nonadditive in a
92 grassland manipulation experiment (Mueller et al., 2016), which agrees with a meta-analysis

93 across different ecosystem types (Dieleman et al., 2012). However, findings from other recent
94 syntheses showed how eCO₂, warming, N addition, P addition, and altered rainfall regimes
95 generally result in additive interactions between individual drivers for plant N concentrations, plant
96 C:N:P stoichiometry, and terrestrial C and P pools (Yuan & Chen, 2015; Yue, Fornara, Yang, Peng,
97 Li, et al., 2017; Yue, Fornara, Yang, Peng, Peng, et al., 2017; Yue et al., 2018). These studies further
98 suggested that the interactive effects of multiple environmental change drivers may vary
99 substantially depending on the tested variables and combinations of drivers.

100 Here, we first compiled data from 760 published articles reporting field manipulative
101 experiments in both natural and managed ecosystems and then adopted a meta-analysis approach
102 to explicitly compare effect sizes of individual, combined and interactive effects of eCO₂, warming,
103 N addition, P addition, rainfall⁺ and drought on plant N, soil N [including total N, inorganic N,
104 ammonium (NH₄⁺), and nitrate (NO₃⁻)] and soil microbial biomass N (MBN) concentrations and
105 pools. Finally, we used structural equation models (SEMs) to assess the role of several
106 environmental and experimental factors (i.e., moderator variables) in influencing the individual,
107 combined, and interactive effects of different environmental change drivers. The main objectives
108 of this study were to (1) quantify the individual and combined effects of multiple drivers on N
109 concentrations and pools in different ecosystem compartments (plants, soils and microbial
110 biomass), (2) assess whether the individual or combined effects of multiple drivers on N
111 concentrations and pools differ, and (3) evaluate whether the interactive effects of these drivers on
112 N concentrations or pools are additive.

113

114 **2 Materials and methods**

115 **2.1 Data extraction and compilation**

116 The *ISI Web of Science*, *PubMed*, and *Google Scholar* were used to search for peer-reviewed
117 journal articles published before 31 January 2018. We focused on field manipulative studies that
118 included the key environmental change drivers eCO₂, warming, N addition, P addition, rainfall⁺,
119 and drought as well as any combination of these six drivers. The criteria for inclusion in our
120 database included the following: (1) manipulative experiments were conducted in the field to
121 collect data regarding at least one of the studied drivers; (2) experimental and control plots were
122 established within the same ecosystem and contrasted in terms of only the target variable; (3) the
123 magnitude of the treatment and the study duration were clearly recorded, and measurements of the
124 variables in the experimental and control groups were performed at the same spatial and temporal
125 scales; (4) the duration of the manipulative experiments was no less than one growing season; and
126 (5) the means, sample sizes, and standard deviations (SD) or standard errors (SE) of the chosen
127 variables were directly provided or could be estimated from the reported data.

128 Plant N concentrations at both the species and community levels were directly recorded from
129 the primary studies, and N pools were either directly recorded or calculated as the product of N
130 concentration and plant biomass. Because most of the primary studies reported data from only
131 mineral soil layers, we considered only the mineral soil layer in this meta-analysis. Data for soil
132 and microbial biomass N concentrations were extracted from the primary studies, while N pools
133 were directly extracted or determined based on the soil bulk density (if available), microbial
134 biomass, and the corresponding N concentrations. In addition, plant biomass data at both the
135 species and community levels were collected only when simultaneous data for N concentrations
136 or pools were reported. When several measurements were taken at different times in a single
137 primary study, we used values from the last measurement to meet the statistical assumption of
138 independence among observations in the meta-analysis (Hedges, Gurevitch, & Curtis, 1999).

139 Furthermore, as the observations from a single primary study representing different plant
140 compartments, ecosystem types and/or climates may not be independent, we used mixed-effects
141 models and treated studies as random factors in the analysis (Koricheva, Gurevitch, & Mengersen,
142 2013). Because the number of studies assessing the combined effects of three or more drivers was
143 too small for a meta-analysis, we considered only driver pairs in this study. When the data in
144 primary studies were presented graphically, the figures were digitized to extract the numerical
145 values using the free software Engauge Digitizer, version 5.1 (Free Software Foundation, Inc.,
146 Boston, MA, USA). Climate variables [i.e., mean annual temperature (MAT) and mean annual
147 precipitation (MAP)] were obtained directly from the primary studies or extracted from *WorldClim*
148 version 2.0 (<http://www.worldclim.org>) using location information in the cases in which these data
149 had not been reported.

150 After extraction, the data from 760 published articles, representing 7622 observations from
151 all continents except Antarctica, were included in our database (see Text S1, Table S1, and Fig. S1
152 in Appendix 1). In this meta-analysis, we investigated the individual effects of each of the studied
153 environmental change drivers, the combined effects of each driver pair (denoted driver 1 + driver
154 2) and the interactions between each driver pair (denoted driver 1 \times driver 2, see Fig. 1a, b). To
155 calculate the first two types of effects, we used the natural-log response ratio (lnRR). We used
156 lnRR because it shows the least bias of the commonly used effect size metrics and its sampling
157 distribution approximates normality (Hedges et al., 1999). In addition, as we want to know the
158 proportional changes in the response variables relative to controls, lnRR is easily interpretable.
159 Interactive effects between two drivers could be calculated from only studies with a full factorial
160 design (see Fig. 1a). Given that few studies met this criterion, we used Hedges' *d* to assess
161 interaction effects because it is an estimate of the standardized mean difference that is not biased

162 by small sample sizes (Gurevitch & Hedges, 2001).

163

164 **2.2 Analysis of individual and combined effects**

165 The effect of one environmental change driver or the combined effect of a driver pair was defined
166 as the response of a variable (e.g., soil N concentration) in a treated sample compared with the
167 value of that variable in the corresponding control (Yue, Fornara, Yang, Peng, Peng, et al., 2017),
168 and was described by lnRR (Hedges et al. 1999). The calculations of lnRR, the associated variance
169 (v_1) and weight (w_1) of each lnRR, and the weighted mean lnRR (lnRR₊₊) are described in detail
170 in Text S2 of Appendix 1. The individual or combined effect was not significant ($P < 0.05$) if the
171 95% confidence interval (CI) of lnRR₊₊ overlapped with zero (Rosenberg, Adams, & Gurevitch,
172 2000). We used the equation $(e^{\ln RR_{++}} - 1) \times 100\%$ to calculate the net responses of N
173 concentrations or pools to the individual or combined effects in terms of the mean percentage of
174 change relative to the control value (%), and the effects were considered not significant at the $p <$
175 0.05 level if the 95% CI overlapped with zero. The lnRR₊₊ and associated 95% CI values were
176 calculated using mixed-effect models in MetaWin 2.1 (Rosenberg et al., 2000).

177 Several environmental and experimental factors may influence the individual, combined, and
178 interactive effects on terrestrial N and were thus included in the meta-analysis as moderator
179 variables (Koricheva et al., 2013). We categorized the constructed database into different
180 subgroups according to ecosystem type (boreal forest, temperate forest, subtropical and tropical
181 forest, grassland, wetland, tundra, shrubland, desert, and cropland), plant functional type (woody
182 and herbaceous), treatment magnitude, type of manipulative facility [open-top chamber (OTC),
183 free-air CO₂ enrichment (FACE), and screen-aided CO₂ control (SACC) for eCO₂; OTC and heater
184 for warming], and fertilizer chemical form (NH₄NO₃, NH₄, NO₃, urea, and mixture of NH₄NO₃

185 and urea) to assess the influence of these categorical moderator variables on effect size. The effect
186 of each categorical moderator on lnRR was evaluated by comparing the heterogeneity within (Q_w)
187 and between (Q_b) moderator levels using mixed-effect models in MetaWin 2.1 (Borenstein,
188 Hedges, Higgins, & Rothstein, 2009). To assess the influence of the continuous moderator
189 variables latitude, MAT, MAP, treatment magnitude, and study duration, an *a priori* conceptual
190 SEM (Fig. 1c) was developed based on current ecological knowledge (Grace, 2006). This SEM
191 was tested separately for each driver or driver pair. We examined the distributions of the
192 endogenous and exogenous variables of the SEM analysis, tested their normality and transformed
193 them when necessary. The covariance between duration and magnitude was included in the model.
194 The analysis was conducted only when the number of data points was > 30 , and we used a
195 bootstrapping method for resampling based on 5000 iterations when the number of data points was
196 < 100 (Grace, 2006). The overall goodness-of-fit of the models was tested using the traditional χ^2
197 goodness-of-fit test and the root mean square error of approximation (RMSEA) index (Grace, 2006;
198 Schermelleh-Engel, Moosbrugger, & Müller, 2003). The SEM analyses were performed with
199 AMOS software version 22.0 (Amos Development Co.).

200

201 **2.3 Analysis of interactive effects**

202 To further assess whether interactive effects were additive, we employed Hedges' d according to
203 established methods (Gurevitch et al., 2000). Accordingly, the interaction effect size (d_I) between
204 drivers A and B was calculated by Eqn (1):

$$205 \quad d_I = \frac{(X_{AB} - X_A) - (X_B - X_C)}{s} J(m) \quad (1)$$

206 where X_C , X_A , X_B , and X_{AB} are the means of a variable in the control, the treatment groups A and
207 B, and their combination group (AB), respectively. The variables s and $J(m)$ are the pooled SD and

208 a correction term for small sample sizes, respectively, which were calculated by Eqns (2) and (3).

$$209 \quad s = \sqrt{\frac{(n_c - 1)s_c^2 + (n_A - 1)s_A^2 + (n_B - 1)s_B^2 + (n_{AB} - 1)s_{AB}^2}{n_c + n_A + n_B + n_{AB} - 4}} \quad (2)$$

$$210 \quad J(m) = 1 - \frac{3}{4m - 1} \quad (3)$$

211 where n_c , n_A , n_B , and n_{AB} are the corresponding sample sizes; s_c , s_A , s_B , and s_{AB} are the SDs in the
 212 control and experimental groups of A, B, and their combination (AB), respectively; and m is the
 213 degree of freedom ($m = n_c + n_A + n_B + n_{AB} - 4$). The variance in d_I (v) was estimated by Eqn (4):

$$214 \quad v = \frac{1}{n_c} + \frac{1}{n_A} + \frac{1}{n_B} + \frac{1}{n_{AB}} + \frac{d_I^2}{2(n_c + n_A + n_B + n_{AB})} \quad (4)$$

215 The weighted mean d_I (d_{++}) was calculated according to Eqn (5):

$$216 \quad d_{++} = \frac{\sum_{i=1}^l \sum_{j=1}^k w_{ij} d_{ij}}{\sum_{i=1}^l \sum_{j=1}^k w_{ij}} \quad (5)$$

217 where l is the number of groups, k is the number of comparisons in the i^{th} group, w is the study
 218 weight [which is also the reciprocal of the variances ($1/v$)] and d is the size of the individual effect.
 219 The 95% CI of d_{++} was calculated as $d_{++} \pm C_{\alpha/2} \times s(d_{++})$, where $C_{\alpha/2}$ is the two-tailed critical value
 220 of the standard normal distribution. When the number of data points was < 20 , we used a
 221 bootstrapping method for resampling to obtain the 2.5% lowest and highest values as CIs based
 222 on 5000 iterations (Janssens et al., 2010; Zhou et al., 2016). The interactions between two drivers
 223 were classified as additive, synergistic, or antagonistic (Jessica Gurevitch et al., 2000; Zhou et al.,
 224 2016). An interactive effect was considered additive if the 95% CI overlapped with zero. If the
 225 individual effects of driver pairs were either both negative or have opposite directions, the
 226 interactions whose total effects were less than 0 were synergistic, and those whose total effects
 227 were greater than 0 were antagonistic. When the individual effects were both positive, the

228 interactions were interpreted conversely (i.e., those >0 were synergistic, and those <0 were
229 antagonistic).

230

231 **3 Results**

232 **3.1 Individual effects of environmental change drivers on terrestrial N concentrations and** 233 **pools**

234 The responses of plant N concentrations and pools to individual drivers differed significantly ($p <$
235 0.001) at both species and community levels (Fig. 2a). Specifically, eCO₂ significantly decreased
236 plant N concentrations by 7% at both species and community levels, but had no effect or a
237 significantly positive (6%) effect on plant N pools at species and community levels. Warming
238 significantly increased plant N pool at species level by an average of 11%, and N addition
239 stimulated plant N concentrations and pools at both species and community levels. P addition
240 significantly increased plant N concentration by 6% at the community level and plant N pools at
241 species (19%) and community levels (24%). Rainfall⁺ showed significant effects on plant N
242 concentration (5%) at the species level. Drought had no effect on plant N concentration or N pools.

243 The responses of plant biomass to individual drivers were generally significantly correlated with
244 the responses of the corresponding plant N pools at both species and community levels, but were
245 not correlated with plant N concentration (Table 1). Soil N concentration significantly increased
246 by 10% and 6% under N and P additions, respectively, while soil N pool was significantly
247 enhanced only by N addition (12%, Fig. 2b). The concentration and pool of soil inorganic nitrogen
248 (SIN) significantly decreased by 11% and 34%, respectively, under eCO₂, but significantly
249 increased by 14% and 19% under warming and by 79% and 125% under N addition. The SIN, soil
250 NH₄⁺ and soil NO₃⁻ were rarely affected by the individual drivers (Fig. S2). In addition, MBN

251 concentration significantly increased by 8% under eCO₂, while MBN pool increased by 17% under
252 N addition.

253

254 **3.2 Combined and interactive effects of environmental change drivers on terrestrial N** 255 **concentrations and pools**

256 The combined effects of N addition + eCO₂, N addition + warming, N addition + P addition, and
257 N addition + rainfall⁺ significantly increased plant N concentration at species level by 4%, 50%,
258 13% and 29%, respectively. Their effects on the corresponding N pool were significantly higher,
259 with increases of 44%, 223%, 89%, and 226%, respectively (Fig. 3a). In contrast, eCO₂ + warming
260 significantly decreased plant N concentration at the species level by 10%, but had no effects on
261 plant N pool at species level. Both plant N concentration and N pool at community level were
262 stimulated significantly by N addition + P addition, with an average increase of 16% and 66%,
263 respectively. In addition, the responses of plant biomass and N pool to the combined effects of
264 environmental change drivers showed significant positive correlations (Table 1).

265 There was no significant difference between N concentration and N pool in relation to the
266 combined effects of environmental change drivers for soil or microbial biomass (Fig. 3b). Soil
267 total N concentration significantly increased by 6% under warming + N addition and N addition +
268 P addition, while soil total N pool was significantly stimulated (12%) only by N addition + P
269 addition. The combined drivers eCO₂ + N addition, warming + N addition, warming + drought,
270 and N addition + rainfall⁺ significantly increased SIN by 36%, 97%, 49%, and 61%, respectively,
271 but had no effect on SIN pool. Soil NO₃⁻ concentration and pool responded similarly to the SIN,
272 while NH₄⁺ concentration was enhanced only by eCO₂ + N addition (+39%) or N addition + P
273 addition (+45%) (Fig. S3). However, neither the concentration nor the pool of MBN was affected

274 by the combined effects (Fig. 3c).

275 In terms of the interactive effects, additive effects on terrestrial N concentrations and pools
276 were more frequently found than synergistic and antagonistic effects across all the driver pairs
277 tested in our study (Fig. 4). With the exception of an antagonistic effect of N addition \times rainfall⁺,
278 the interactive effects of other driver pairs on plant N concentration at the species level were all
279 additive (Fig. 4a). The mean interactive effect of N addition \times P addition on plant N concentration
280 at the community level was antagonistic, but their effect on plant N pool at both the species and
281 community level was synergistic. All interactive effects on the concentrations and pools of soil
282 total N and SIN were additive except for that of N addition \times P addition on SIN concentration,
283 which was synergistic. Likewise, similar patterns were found for NH₄⁺ and NO₃⁻ (Fig. S4), and the
284 interactive effects of all driver pairs on MBN concentrations and pools were additive (Fig. 4).
285 Moreover, despite the observation of several overall non-additive interactive effects, the frequency
286 distribution of interaction types indicates that additive interactions were substantially predominant
287 across all the driver pairs (Fig. 4).

288

289 **3.3 Influences of moderator variables**

290 Moderator variables such as ecosystem type, experimental design factors (e.g., study duration,
291 facility, treatment magnitude, and fertilizer form), latitude, and climate (i.e., MAT and MAP) all
292 mediated the responses of terrestrial N concentrations and pools to individual, combined, and
293 interactive effects of the investigated drivers (Fig. 5 and Fig. S5-S13 in Appendix 2). For example,
294 the individual effects of eCO₂ on plant N concentration ($Q_b = 15.25$, $P = 0.033$) and pool ($Q_b =$
295 18.50 , $P = 0.010$) varied significantly with ecosystem type (Fig. S5a) and were also significantly
296 influenced by the magnitudes of eCO₂, MAP, and study duration (Fig. S13a). Likewise, N addition

297 + P addition on plant N concentration and pool significantly varied with ecosystem type (Fig. S18c,
298 d) and were significantly influenced by latitude, MAT, and MAP in some cases (Fig. 5a), although
299 such impacts on N concentrations and pools vary with terrestrial compartment. The effect of eCO₂
300 × warming on plant N concentrations was significantly influenced by latitude and MAP, while N
301 addition × P addition on plant N pools at both species and community levels were significantly
302 modulated by experiment duration (Fig. 5b).

303

304 **4 Discussion**

305 Our results show how N concentrations and N pools in plants, soils, and microbial biomass are
306 influenced by the individual and combined effects of different environmental change drivers. We
307 found that N concentrations and pools in plants show higher sensitivity to these driver effects than
308 those in soils and microbial biomass. The different responses between plant N concentrations and
309 N pools to the individual or combined effects of multiple drivers suggest that mixing plant N
310 concentration and N pool data can be problematic when plant N cycling is assessed under different
311 environments. However, such a difference was seldom observed in soils or microbial biomass. In
312 addition, our results show that the interactive effects of multiple drivers on N concentrations and
313 N pools of plants, soils and microbes are more likely to be additive. These novel findings contribute
314 to improving our understanding of how terrestrial N cycling among plants, soils, and microbes
315 may shift under the simultaneous effects of multiple environmental change drivers.

316

317 **4.1 Differential individual effects of environmental change drivers on terrestrial N** 318 **concentrations and pools**

319 The magnitude and direction of environmental change effects on terrestrial N vary depending on

320 the identity of the driver and the nature of the N concentration or pool (i.e., plants *vs.* soils *vs.* soil
321 microbial biomass). For example, our results show how eCO₂ significantly decreased plant N
322 concentration at both species and community levels, but significantly increased plant N pool at the
323 community level (Fig. 2a). Net negative effects of eCO₂ on plant N concentration may occur
324 because of (1) dilution of N by increased photosynthetic assimilation of C; (2) lower transpiration-
325 driven mass flow of N in soils due to decreased stomatal conductance under eCO₂; and (3)
326 increased rates of N loss via volatilization and/or root exudation, which may further decrease tissue
327 N concentration (Taub & Wang, 2008). The increase in plant N pools under eCO₂ could be related
328 to eCO₂-induced increases in biological N fixation or increases in root growth for N uptake (Luo
329 et al., 2004). Warming-induced net N accumulation in plants (i.e., plant N pool) may be attributed
330 to enhanced plant growth that is being stimulated by changes in soil N availability (Vitousek &
331 Howarth, 1991) and in plant phenology (Luo, Sherry, Zhou, & Wan, 2009) under warming
332 conditions. This idea was supported by our results showing that SIN concentration and pool
333 significantly increased in the warming treatments (see Fig. 2a).

334 Nitrogen addition generally showed positive effects on plant N concentrations and N pools
335 (Fig. 2a) and on soil N pools (Fig. 2b). These findings agree with the results of previous meta-
336 analyses, which showed that N addition had positive effects on plant N concentrations and N pools
337 (Bai et al., 2013; Lu et al., 2011). The significant positive effects that P addition and rainfall⁺ had
338 on plant N concentrations and pools (Fig. 2a) may be attributed to the fact that both P and water
339 availability are limiting factors for plant growth and N uptake (Li et al., 2016). In contrast, drought
340 showed minimal effects on plant, soil and microbial biomass N concentrations and pools.
341 Microbial biomass N pools were significantly increased by eCO₂ (Fig. 2c). Previous studies
342 showed that soil microbial numbers, metabolic activity, and biomass can be increased by eCO₂

343 (Sadowsky & Schortemeyer, 1997). Thus, this increase can be attributed to eCO₂ increasing the
344 microbial utilization of soil organic matter (Carney, Hungate, Drake, & Megonigal, 2007) and N
345 fixation ability through stimulating the activities of related enzymes (Cheng et al., 2011; He et al.,
346 2010).

347 Our results show that the responses of plant N concentrations and N pools to the individual
348 effects of multiple drivers can vary significantly (Fig. 2a), suggesting that the effects on these two
349 variables should be separately tested when plant N dynamics are assessed in future studies. We
350 found that while the responses of plant biomass to these drivers were significantly correlated with
351 the corresponding plant N pools, they were not related to plant N concentrations (Table 1). This
352 phenomenon is likely to occur when one driver or driver pair increases plant biomass without
353 changing N concentration, thus resulting in an increase in the N pool (Doiron, Gauthier, &
354 Lévesque, 2014). Our evidence is that the responses of soil N and MBN to the individual drivers
355 are weaker than plant N responses (Fig. 2), which may partly occur because soil N pools are much
356 larger than plant N pools (Nieder & Benbi, 2008) and thus expected to respond more slowly to
357 environmental change. In addition, changes in soil N pools may also be difficult to detect because
358 of the higher inherent variation in soil organic matter under environmental change.

359

360 **4.2 Combined effects and the interactions of driver pairs**

361 Our results showed that eCO₂ + N addition, eCO₂ + drought, and warming + N addition stimulated
362 plant N concentrations and N pools at the species level (Fig. 3a). The positive effects of eCO₂ + N
363 addition on plant N concentrations and N pools may be attributed to the fact that additional N input
364 meets the increased plant N demand under eCO₂ (Reich, Hobbie, & Lee, 2014). The stimulating
365 effects of eCO₂ + drought on plant N pools were mainly attributed to the net positive effect of

366 eCO₂ because, as shown in our results, drought had no effect on plant N pools and the interactive
367 effect of eCO₂ × drought was additive.

368 Plant growth and plant biomass production in terrestrial ecosystems are primarily limited by
369 N availability or are co-limited by P availability (Elser et al., 2007; Vitousek, Porder, Houlton, &
370 Chadwick, 2010). Larger inputs of one nutrient (either N or P) will stimulate plant growth but also
371 lead to increased demand for the other nutrient. Thus, the simultaneous addition of N and P can
372 significantly contribute to increasing plant N uptake rates, which can lead to increased N
373 concentrations and N pools (Li et al., 2016). N addition × P addition on plant N pools was indeed
374 synergistic at both species and community levels, albeit antagonistic for plant N concentrations at
375 community level (Fig. 4). N addition increases the demand for P by stimulating plant growth; thus,
376 extra P addition could counterbalance this N-induced P limitation, allowing the full N fertilization
377 effect to be expressed (You et al., 2018). In this case, the N addition effect could be larger in the
378 combined treatment with P than in the N-only treatment, resulting in a synergistic N addition × P
379 addition effect. However, the occurrence of such a synergistic effect is conditional, as the
380 interactive effect was significantly influenced by experimental and environmental factors such as
381 latitude and study duration (see Fig. 5b). The full expression of the N addition effect with
382 additional P input can be conditional because net N addition effects on plant N uptake are soil
383 dependent and thus are influenced by multiple biogeochemical factors (Niu et al., 2016). When
384 increases in plant biomass are larger than increases in N uptake, N addition × P addition effects on
385 plant N concentration could be antagonistic but could become synergistic for N pools as a result
386 of increases in N mass. However, despite these non-additive effects, our results indicated that
387 additive effects were much more common across individual observations and that the overall effect
388 of N addition × P addition on plant N concentrations at the species level was additive.

389 We found that N addition + rainfall⁺ had significant positive effects on plant N concentrations
390 and N pools at the species level (Fig. 3a) and that the mean interactive effect of N addition ×
391 rainfall⁺ on plant N concentrations was antagonistic (Fig. 4). Water availability is an important
392 factor regulating ecosystem primary production, and increased water input generally stimulates
393 biomass production through increased uptake of limiting nutrients such as N (Li, Lin, Taube, Pan,
394 & Dittert, 2011). However, although N addition can enhance plant N uptake, it can also decrease
395 plant N use efficiency (Lü, Dijkstra, Kong, Wang, & Han, 2014); thus, the interaction of N with
396 rainfall⁺ can be antagonistic for plant N concentrations. Nevertheless, the additive interactions of
397 N addition × rainfall⁺ on plant N concentrations remained predominant at the species level.

398 Soil total N concentration was significantly enhanced by warming + N addition and N
399 addition + P addition (Fig. 3b). Although N addition alone significantly increased soil N
400 concentration, further positive effects on soil total N occurred when the N fertilization effect (i.e.,
401 N-induced increases in plant N input to soils) was enhanced by warming or P addition (Bai et al.,
402 2013). Because the SIN pool is smaller than the soil total N pool (Benbi & Richter, 2003), the
403 variations in SIN can be more sensitive than the variations in the soil total N pool to environmental
404 change drivers, as we found in this study (Fig. 3b). Common additive interactions were also
405 observed for soil total N and SIN, although we observed an overall synergistic effect of N addition
406 × P addition on SIN concentration (Fig. 4). N addition can increase phosphatase activity and thus
407 soil P availability (Olander & Vitousek, 2000), but this N-induced potential increase in P
408 availability is usually insufficient to balance the accompanying increased P limitation. Thus, the
409 simultaneous addition of P and N would enhance SIN because P addition can also stimulate rates
410 of N fixation and increase N availability (Crews, Farrington, & Vitousek, 2000). Hence, the
411 synergistic effects of N addition × P addition on SIN concentration are not unrealistic. In addition,

412 the overall synergistic effect on SIN could result from the large weight that individual synergistic
413 observations may have relative to more common ($> 50\%$) additive effects (Zhou et al., 2016). We
414 observed similar patterns of additive interactive effects on MBN concentrations. However, in
415 contrast to plant and soil N, soil MBN exhibited no significant response to the combined effects
416 of driver pairs, which may be attributed to the small sample size that limited the breadth of our
417 analysis (Loladze, 2014).

418

419 **4.3 Moderating effects of environmental and experimental factors**

420 Moderator variables, such as environmental and experimental factors, which affect the individual
421 effects of environmental change drivers on specific ecosystem properties, have been highlighted
422 and discussed in previous studies (Bai et al., 2013; Li et al., 2016; Lu et al., 2011; Xia & Wan,
423 2008). Similarly, our results revealed that such moderator variables significantly influence the
424 combined and interactive effects of driver pairs on plant and soil N concentrations and pools. For
425 example, the combined effects of $e\text{CO}_2 + \text{N}$ addition on plant N concentrations and pools
426 significantly varied with ecosystem type, with significantly positive effects in grassland, but no
427 effects in tropical forests. This difference may be because $e\text{CO}_2$ -induced nutrient limitations were
428 related to not only N, but also other nutrients such as P (Winter, Garcia, Gottsberger, & Popp,
429 2001), whose availability is typically limited in tropical forests (Elser et al., 2007). The interactive
430 effects of $e\text{CO}_2 \times \text{N}$ addition on plant N were also significantly affected by moderator variables
431 such as latitude and MAP (Fig. 5b). As discussed above, driver effects are expected to be
432 environmentally dependent. Confirming this expectation, our results suggest that MAP influences
433 the extent to which driver effects are expressed and modulates the interactive effects of driver pairs
434 on N cycling. Likewise, we found that the effects of N addition + P addition and N addition \times P

435 addition on the concentrations and pools of plant and soil N were significantly modulated by
436 latitude, MAT, MAP, and experimental duration (Fig. 5b, S13). This relationship could be
437 attributed to ecosystem N cycling differing along gradients generated by these environmental and
438 experimental factors (Bai et al., 2013; Lu et al., 2011). Moreover, the effect size of N addition + P
439 addition on soil NO_3^- concentration was significantly positively correlated with the duration of the
440 manipulation study (Fig. S13), indicating that long studies are necessary to completely assess
441 environmental change effects on terrestrial N pools.

442

443 **4.4 Uncertainty analysis and limitations**

444 Although our meta-analysis provides new evidence of how the individual, combined, and
445 interactive effects of multiple environmental change drivers might influence terrestrial N
446 concentrations and N pools, significant uncertainty still remains. First, primary studies assessing
447 the effects of environmental changes on plants, soils, and microbes are not abundant enough and
448 unequally distributed geographically, thus limiting our ability to conduct a global analysis of N
449 cycling among different ecosystem components. Evaluating combined effects is even more
450 challenging because the small sample sizes of data for many driver pairs hampered our ability to
451 quantify the nature of the interactions. Additionally, the lack of data for other driver pairs and
452 potential combinations of three or more drivers reduced the breadth of our analysis. Second,
453 published studies on global change effects on nitrogen dynamics are strongly biased towards the
454 northern hemisphere (see Fig. S1); thus, the database failed to represent ecosystem types equally
455 at the global scale, especially savanna and tropical forest ecosystems. Third, in terms of statistics,
456 the observation-weighted approach that we used here might overestimate the amount of additive
457 interactions associated with large variance in some observations (Gurevitch et al., 2000; Zhou et

458 al., 2016). Nevertheless, our statistical analysis showed that the average weights of the interaction
459 (*d*) for significant results (synergistic and antagonistic effects) were similar to those for the non-
460 significant interactions (additive effects) (see Table S3). This result suggests that overestimation
461 of additive interactions should not be an issue (Zhou et al., 2016).

462

463 **4.5 Implications and perspectives**

464 Our study shows how plant N concentrations and N pools were differently affected by the
465 individual and combined effects of multiple environmental change drivers, indicating that merging
466 these two N variables would potentially be misleading when assessing the response of plant N
467 dynamics to environmental change. The weak responses of soil and microbial N concentrations
468 and N pools (compared to plants) to both individual and combined effects of multiple drivers
469 suggest that soils and soil microbes are less sensitive than plants to environmental change or that
470 their responses are more difficult to detect. Most importantly, our study suggests that responses of
471 N concentrations and N pools in different terrestrial compartments to the interactive effects of
472 multiple drivers are more likely to be additive than synergistic or antagonistic. These common
473 additive effects of driver pairs on N concentrations and pools should be incorporated into
474 ecosystem models that aim to predict how altered N availability affects global C sinks. Future
475 studies could address what underlying biogeochemical mechanisms enhance the stability of soils
476 and soil microbes to environmental change. Finally, well-designed long-term experiments that
477 simultaneously assess effects of multiple drivers on ecosystem compartments are urgently needed
478 to better capture the dynamics of terrestrial N cycling and their consequences for terrestrial C
479 storage under future environmental change scenarios.

480

481 **Data accessibility**

482 All the data used in the meta-analysis are included in the Supporting Information.

483

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619

620 **Supporting Information**

621 **Appendix 1** Data sources and methods.

622 **Appendix 2** Supporting results.

623

624

625 **Table 1** Pearson correlation coefficients (r) between the effect sizes of global change driver (lnRR) on plant
 626 biomass and terrestrial N concentrations (-C) and pools (-P).

Terrestrial N	eCO ₂	Warming	N addition	P addition	Rainfall [†]	Drought	eCO ₂ + N	N + P	N + rainfall [†]	eCO ₂ + warming
PN-C (species)	-0.102 (136)	0.152 (46)	0.054 (203)	-0.104 (51)	-0.085 (11)	0.184 (10)	0.372 (15)	-0.255 (25)	-0.028 (9)	-0.476 (13)
PN-P (species)		0.801 ^{***} (43)	0.925 ^{***} (216)	0.942 ^{***} (45)	0.907 ^{***} (11)	0.697 [*] (10)	0.989 ^{***} (15)	0.960 ^{***} (35)	0.969 ^{***} (9)	0.609 [*] (9)
PN-C (community)	-0.405 (11)	0.002 (8)	-0.254 [*] (92)	-0.064 (50)				-0.130 (45)		
PN-P (community)	0.971 ^{***} (15)	0.843 ^{***} (15)	0.608 ^{***} (93)	0.873 ^{***} (44)		0.880 ^{***} (12)		0.797 ^{***} (42)		
SN-C		< 0.001 (13)	-0.099 (105)	-0.042 (22)				-0.216 (30)		
SN-P		0.046 (30)	0.078 (56)					-0.036 (14)		
SIN-C		0.227 (11)	0.148 (59)	-0.037 (23)				-0.113 (27)		
SIN-P			0.166 (27)							
NH ₄ ⁺ -C		0.385 (14)	0.205 (205)	0.130 (17)				-0.192 (20)		
NH ₄ ⁺ -P		0.617 (10)	0.609 (8)					-0.327 (9)		
NO ₃ ⁻ -C		0.444 (13)	-0.058 (54)	-0.061 (17)				-0.318 (20)		
NO ₃ ⁻ -P			0.750 [*] (8)					-0.009 (9)		
MBN-C			-0.654 ^{**} (17)							
MBN-P		0.155 (23)								

627 Values in brackets indicate the sample size of observations. Correlation analysis was conducted only when the number of data points was > 8. Asterisks
 628 indicate significant (^{*} $p < 0.05$, ^{**} $p < 0.01$, ^{***} $p < 0.001$) correlations. PN: plant nitrogen; SN: soil total nitrogen; SIN: soil inorganic nitrogen; MBN: microbial
 629 biomass nitrogen; eCO₂: elevated CO₂; rainfall[†]: increased rainfall.

630

631

632

633 **Figure captions**

634 **Figure 1** (a) Three common experimental designs included in our database to study the effects of
635 two environmental change drivers, A and B, with C being the control plot. Individual effects of
636 drivers A and B were calculated based on data from study designs 2 and 3. The combined effects,
637 A+B, were calculated based on data from designs 1 and 3, while the interactive effects, $A \times B$,
638 could be calculated from only study design 3 (i.e., $A+B = A + B + A \times B$). (b) A real picture of the
639 Cedar Creek Ecosystem Science Reserve in Minnesota, US. Several experiments at this research
640 site examined the interactions between multiple environmental drivers [Photo credit: Jacob Miller,
641 2014, CC BY-SA 4.0]. (c) An *a priori* conceptual structural equation model (SEM) depicting the
642 influence of latitude, longitude, mean annual temperature (MAT), mean annual precipitation
643 (MAP), driver magnitude, and study duration on the effect size (lnRR or d_I) of environmental
644 change drivers on terrestrial N concentrations or pools. The same model was used for all tested
645 drivers and driver pairs. Single-headed arrows indicate a hypothesized directional influence of one
646 variable on another, double-headed arrows represent a correlation in which no direction is specified,
647 and each square indicates a measured variable entered in the model. Note that “magnitude” was
648 tested for only individual drivers and the combination N addition + P addition, in which case the
649 ratio between the added N and P (N:P) was used.

650

651 **Figure 2** Individual effects of multiple environmental change drivers on terrestrial N
652 concentrations and pools in (a) plants, (b) soils and (c) soil microbial biomass. The results are
653 expressed as the percentage change relative to the control (%). Values indicate the means with 95%
654 confidence intervals (CIs), and sample size numbers for N concentrations and pools are shown in
655 parentheses. The effects of environmental change drivers are significant when the 95% CIs do not

656 overlap with zero. The results are not presented when the sample size is < 3 . Asterisks indicate
657 significant ($*p < 0.05$, $**p < 0.01$, $***p < 0.001$) differences between the responses of N
658 concentrations and pools to a specific driver. eCO₂: elevated CO₂; N: nitrogen addition; P:
659 phosphorus addition; Rainfall⁺: increased rainfall.

660

661 **Figure 3** Combined effects of multiple environmental change drivers on terrestrial N
662 concentrations and pools in (a) plants, (b) soils and (c) soil microbial biomass. The results are
663 expressed as the percentage change relative to the control (%). Values indicate the means with 95%
664 confidence intervals (CIs), and sample size numbers for N concentrations and pools are shown in
665 parentheses. The effects of environmental change drivers are significant when the 95% CIs do not
666 overlap with zero. The results are not presented when the sample size is < 3 . Asterisks indicate
667 significant ($*p < 0.05$, $**p < 0.01$, $***p < 0.001$) differences between the responses of N
668 concentrations and pools to a specific driver pair. eCO₂: elevated CO₂; N: nitrogen addition; P:
669 phosphorus addition; Rainfall⁺: increased rainfall; MBN: microbial biomass nitrogen.

670

671 **Figure 4** Interactive effects of multiple environmental change drivers on terrestrial (a) N
672 concentrations and (b) N pools and the corresponding frequency distribution of interaction types
673 among individual observations of driver pairs. Dots represent means with 95% confidence
674 intervals (CIs), and sample size numbers are shown in parentheses. If the 95% CI overlapped with
675 zero, the interactive effect was considered to be additive; otherwise, the interactive effect was
676 synergistic or antagonistic. The results are not presented when the sample size is < 3 . Because
677 many studies reported only combined effects, sample sizes may be smaller than the corresponding
678 ones in Fig. 3. Values in percentages indicate the proportions of additive interactions among all the

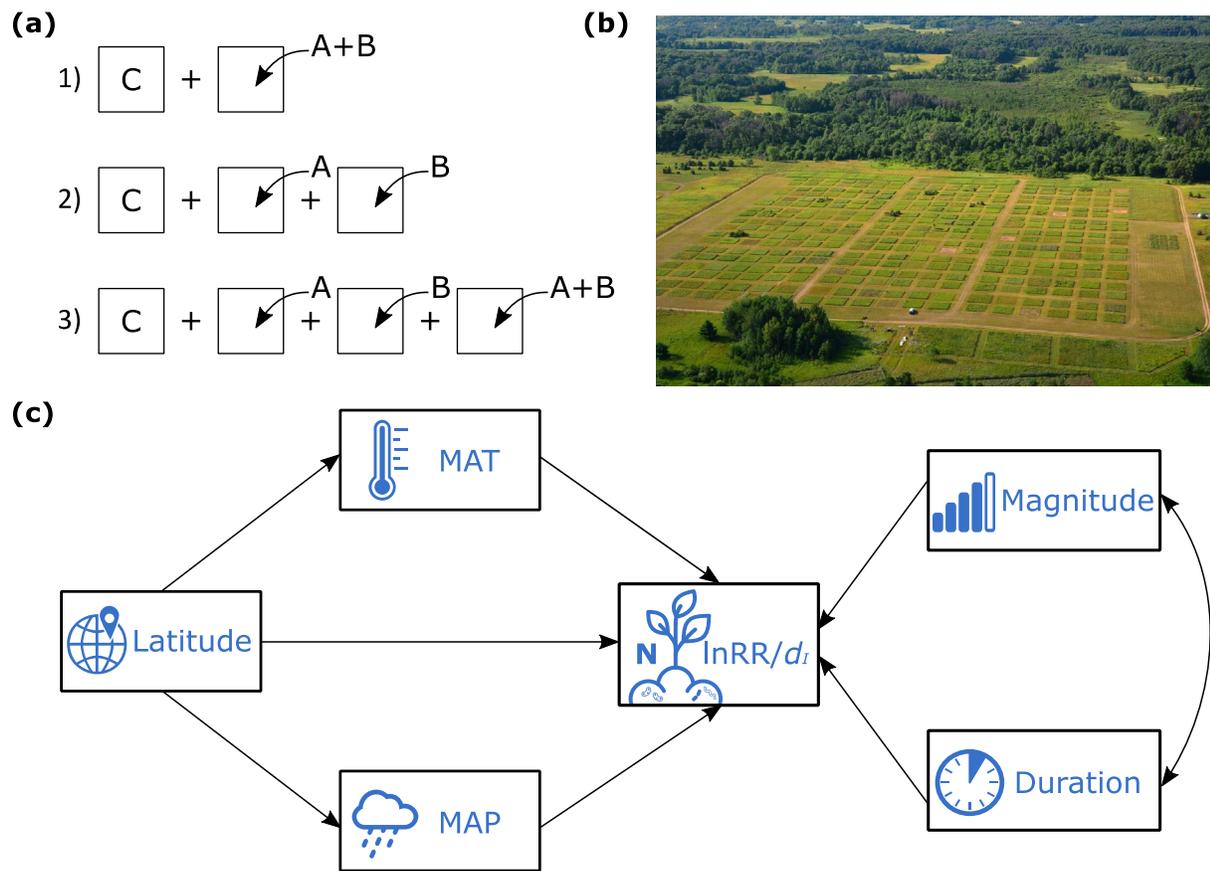
679 individual observations of a specific response variable. PN: plant nitrogen; SN: soil total nitrogen;
680 SIN: soil inorganic nitrogen; MBN: microbial biomass nitrogen; eCO₂: elevated CO₂; N: nitrogen
681 addition; P: phosphorus addition; Rainfall: increased rainfall.

682

683 **Figure 5** Standardized total effects (direct plus indirect effects) derived from structural equation
684 models (SEMs) evaluating the influence of moderator variables (depicted in different colors) on
685 (a) the combined effects (lnRR) and (b) interactive effects (*d_i*) of driver pairs on terrestrial N
686 concentrations (-C) and pools (-P). Note that the magnitude of the tested driver pairs was assessed
687 in only N and P combinations, where the ratio between the added N and P (N:P ratio) is used.
688 Asterisks indicate significant (**p* < 0.05, ***p* < 0.01, ****p* < 0.001) direct effects. See Table S2 in
689 Appendix 2 for the goodness-of-fit tests of the SEMs, and see Fig. S13 in Appendix 2 for the
690 results derived from the SEMs evaluating the influence of moderator variables on the individual
691 effects. eCO₂: elevated CO₂; PN: plant nitrogen; SN: soil total nitrogen; SIN: soil inorganic
692 nitrogen; MBN: microbial biomass nitrogen.

693

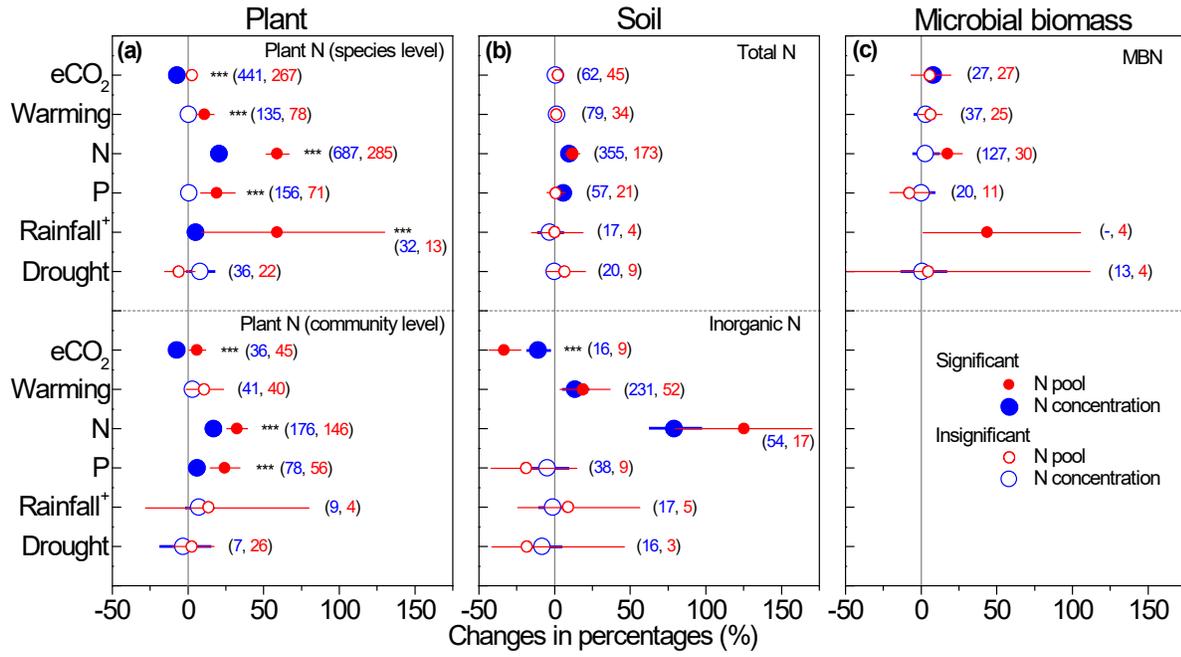
694 **Figure 1**



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697 **Figure 2**

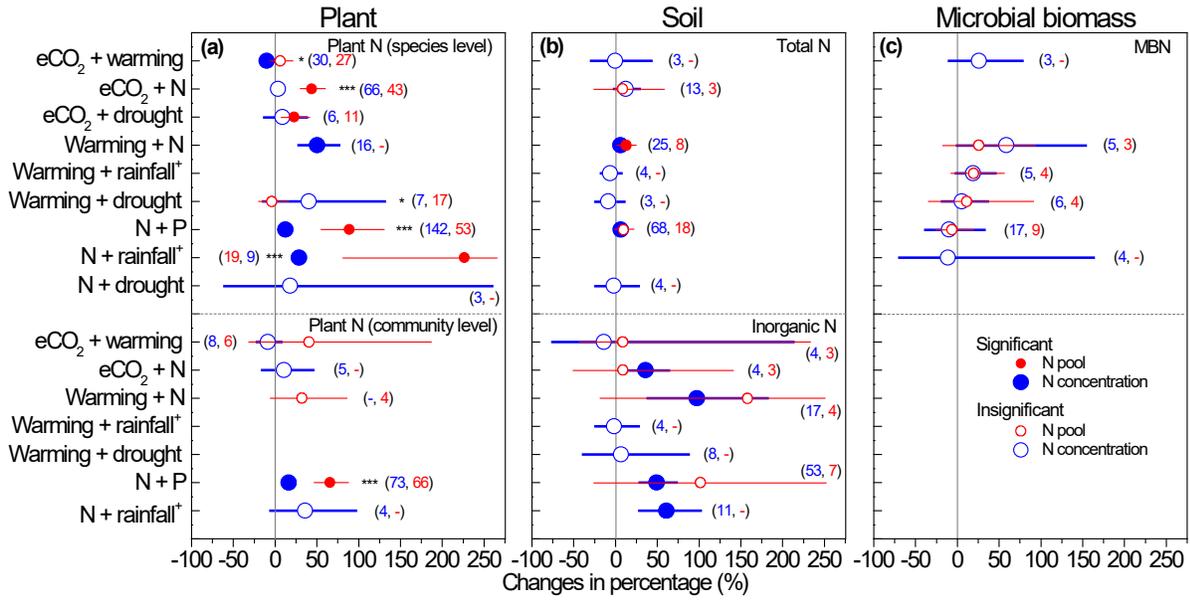


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701 **Figure 3**

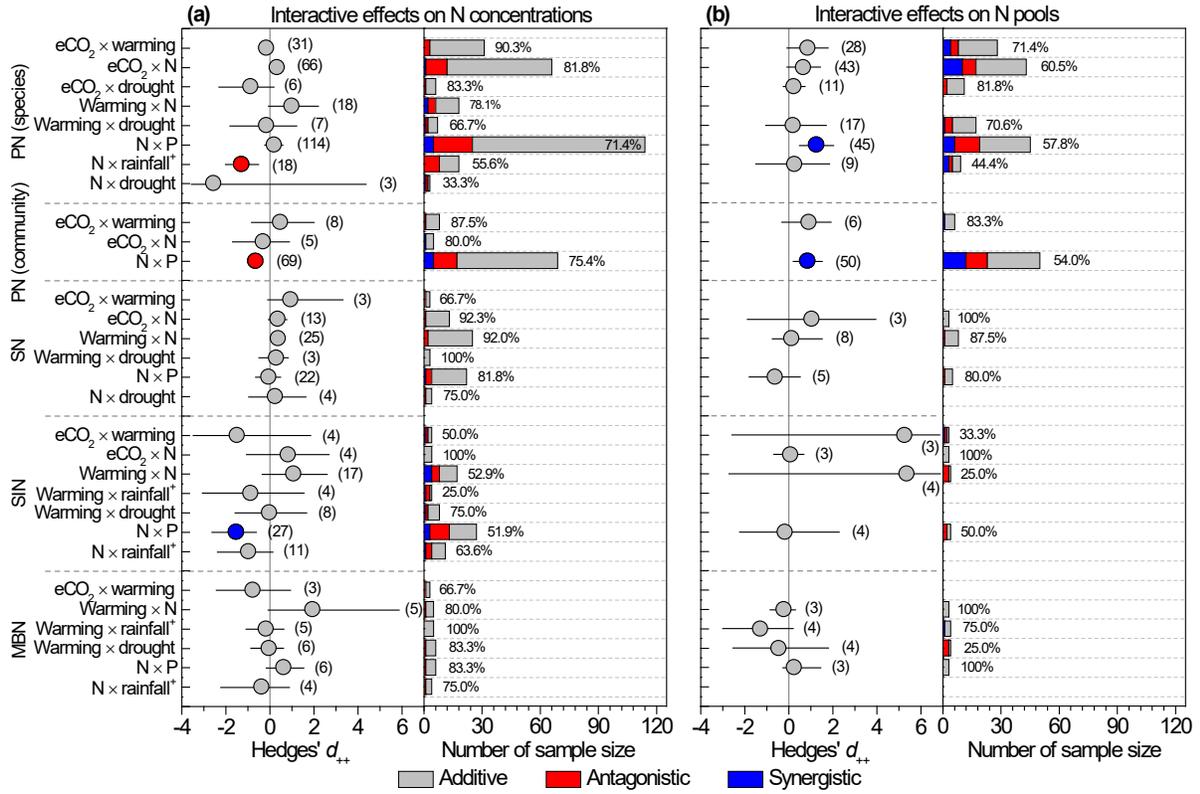


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705 **Figure 4**

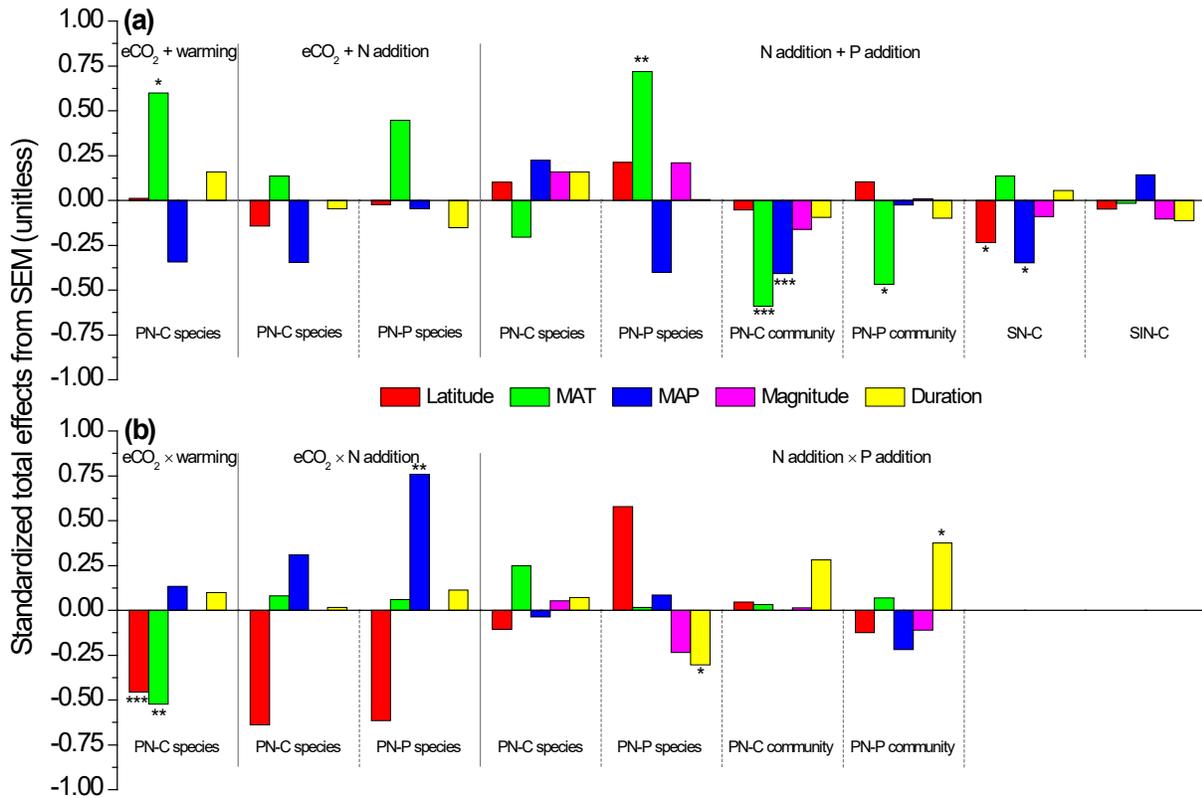


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709 **Figure 5**



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